

**Population size and habitat relationships of Black-bellied
Sandgrouse, *Pterocles orientalis*, in the Canary Islands, Spain.**

Seoane, J.* , Carrascal, L.M.** , Palomino, D*** . and Alonso, C.L**** .

* Dept. Interuniversitario de Ecología, Facultad de Ciencias, Universidad Autónoma de Madrid, 28049 Madrid, Spain

** Dept. Biodiversidad y Biología Evolutiva, Museo Nacional de Ciencias Naturales, CSIC. C/José Gutiérrez Abascal 2, 28006 Madrid, Spain

*** Área de Estudio y Seguimiento de Aves, Sociedad Española de Ornitología (SEO/BirdLife), C/Melquiádes Biencinto 34, 28006 Madrid, Spain

**** Instituto de Ciencias Ambientales (ICAM), Universidad de Castilla-La Mancha, 45071 Toledo, Spain

Author for correspondence: Javier Seoane. E-mail: javier.seoane@uam.es. Phone : +34 91 4973639. Fax : +34 91 4978001.

Summary

We estimate the breeding population size and assess the habitat relationships of Black-Bellied Sandgrouse in the Eastern Canary Islands (Fuerteventura, Lanzarote and La Graciosa, Spain) by means of a survey based on 1,787 0.5 km-line transects and distance sampling done in 2005 and 2006. The population comprises 2,906 individuals (90% CI: 2,363-3,562), which is much higher than the numbers estimated in previous reports based on partial surveys and constitutes 20% of the total Spanish population. Sandgrouses in the Canaries are currently restricted to Fuerteventura, where 70% of the population gathers in four areas that encompass just 16.7 % of the island and are largely within Special Protection Areas classified under Birds Directive (except the area of Tefia-Ampuyenta, first in absolute number of individuals). The environmental characteristics that maximize the probability of occurrence of the sandgrouse in Fuerteventura (probability = 0.196) are: treeless non-cultivated areas of sandy soils without bare bedrocks, with a rock cover less than 44%, located in non-coastal areas with an average terrain slope lower than 27.5%, at more than 400 m from the nearest urban area, with less than 795 m of dirt roads per 20 ha, with at least 0.9% of shrub cover and a NDVI index higher than 53. Sandgrouses were closer to human settlements in midsummer than in March, perhaps being attracted to artificial pools surrounding villages. Similar habitat characteristics exist in nearby Lanzarote, where the species could hypothetically reach densities as high as 4-5 birds / km². Possible reasons for the absence of sandgrouses in this island are discussed.

Running head: Black-Bellied Sandgrouse in the Canaries

Keywords: Eastern Canary Islands, distance sampling, habitat modelling, population density, population size.

Introduction

Open habitats like semi-deserts and grasslands have been extensively modified throughout the world and currently much of their former natural cover has been lost to crop and cattle fields, or urban and industrial developments, with low and relatively flat areas being most affected due to its greater accessibility (Onrubia and Andrés 2005). Tropical and temperate islands are currently in a particular high risk of this form of land transformation due to the high economic pressure to devote the land to recreational facilities (hotels, golf courses) that compete for local and frequently scarce resources, most notably water in semi-desert areas. Consequently, island birds from open-country habitats frequently appear in the lists of endangered species (Groombridge 1992; BirdLife-International 2000).

The Black-bellied sandgrouse *Pterocles orientalis* is a Palearctic species that reaches the western border of its distribution range in the Canary Islands archipelago. Currently, the only breeding populations occur in Fuerteventura, although there are a few records of the species in the close and similar island of Lanzarote (Emmerson 1999; Martín and Lorenzo 2001; Emmerson and Lorenzo 2007). It occupies open arid, semi-desert and steppe habitats where it forages for seeds on the ground in variable-sized flocks, which can be most easily detected when flying to pools and creeks where they gather for watering (De Juana 1997). Western Palearctic populations (*P. o. orientalis*) have been quantified to 62,000 pairs, mainly distributed in the strongholds of Turkey (25,000 to 50,000 pairs; BirdLife-International 2004) and Spain (7,824 to 13,273 pairs; Suárez *et al.* 2006). It is classified as Vulnerable in the Red Book of Spanish Birds because of a decline in extent of occurrences and negative population trends (Suárez and Herranz 2004). In the Canary Islands, however, it has never been adequately researched, and there is an urgent need of an accurate census of its population and

knowledge of its habitat preference to support planning and assessment, given the increasing pressure for land-use transformation in the archipelago (Fernández-Palacios and Martín Esquivel 2001).

In this work we first estimate the population size and identify the more important areas for Black-bellied sandgrouse in the Canary Islands. We also model the habitat preferences of the species in Fuerteventura and apply the results to the nearby island of Lanzarote in order to recognize potential breeding areas and discuss how many birds they could harbour and why they are currently unoccupied.

Methods

Study area

Fuerteventura is an eastern island of the Canary archipelago (the second largest one: 1730 km²; 28°27' N, 14°00' W), lying only 100 km far from the North-African coast. It shows a smooth relief (highest altitude: 807 m) in accordance with its ancient geological origins (20-22 million years) and subsequent erosion, since the volcanic activity of the island is almost extinct. The combined effects of direct Saharan influence on climate and a prevailing flat topography result in a dominance of scarcely vegetated arid landscapes, which have been extensively grazed (mainly by goat herds) and cultivated. The impoverished plant communities mostly consist of a few species of xerophytic shrubs (*Launaea arborescens*, *Lycium intricatum*, *Salsola vermiculata*, *Suaeda* spp. and *Euphorbia* spp.), therophytic forbs and several perennial grass species. The only natural woodlands are small and patchily located tamarisk and palm groves (*Tamarix canariensis* and *Phoenix canariensis*, respectively). The degree of development of vegetated areas is relatively diverse due to local conditions, such as humidity, slope of terrain, soil characteristics, goat grazing, and human uses. With

regard to soil lithology and compactness, the study areas also comprise a broad range of conditions, from stony lava fields to loose sand dunes (Rodríguez *et al.* 2000; Fernández-Palacios and Martín Esquivel 2001). Just 12 km to the north, the close island of Lanzarote shares the volcanic origin, climatic characteristics and many of its general habitat attributes with Fuerteventura, but its smaller (846 km²), shorter (maximum altitude 846 m), younger (16-19 million years) and has a higher proportion of its surface covered by lava fields. The volcanic island of La Graciosa, 2 km northwards Lanzarote, completes our study area with a smaller territory (18.7 km²) mainly occupied by sandy areas and small hills (maximum altitude 266 m) covered by bushes (see Fernández-Palacios and Martín-Esquivel, 2001 for more details).

Bird and habitat data

Breeding bird surveys in the islands were carried out during the periods 12-26/02/2005 (Lanzarote), 22-23/02/2005 (La Graciosa) and 05/03/2005-09/04/2005 and 05/03/2006-14/03/2006 (Fuerteventura). The phenology of the species is little known, but recent accounts mention March to July as the breeding season, so we feel confident the survey covered the early breeding period (Martín and Lorenzo 2001; Emmerson and Lorenzo 2007). Moreover, the average flock size of the species was 2.62 sandgrouses (sd=1.57), the interquartile range was 2-3 birds, and the percentage of flocks with four or less sandgrouses was 90.8% in Fuerteventura during the census period (n=152 near contacts at less than 100 m from the observer, for which it is highly probable that all birds were detected). These figures are nearly identical to those reported for the species during the breeding season in the Iberian peninsula (De Borbón *et al.* 1999: 78-85% of flocks with 1-4 birds in June and July, average of 3.0 birds per flock during the breeding season).

The survey method used was the line transect, frequently used in extensive assessments of abundance, general distribution patterns and habitat preferences of birds, (Bibby *et al.* 2000), which we have previously used to assess land bird populations in the Canary Islands (e.g.: Carrascal *et al.* 2006; Palomino *et al.* 2008). Line transects of 0.5-km (geolocated and measured by means of portable GPSs) were performed across the whole island (Figure 1), including all of the main non-urban habitats: barren lava fields, shrubby steppe-like plains, stony/sandy desert areas, traditional cultivations, hilly/mountain slopes, and gullies. Line transects were carried out on windless and rainless days, walking cross-country or on dirt tracks at a low speed (1-3 km/h approximately), during the first four hours after dawn and the two and a half hours before dusk.

Additionally, to assess seasonal changes in the distribution/abundance of the Black-bellied Sandgrouse, a habitat-stratified selection of 602 transects was repeated in the post-breeding period of 2006 (from 07 to 15/08/2006). The sampling locations and the approximate number of transects to gather on them, were roughly determined in proportion to the surface in the islands of each type of main landscape types. Apart from the mere availability of a safe place to park, the starting point of each sampling line was randomly determined. Next, the observers walked through the target area trying: a) to perform 0.5 km transects as homogeneous as possible; b) to attain an extensive cover of the surveyed area. The transect lines were not biased by an a priori potential of the habitat to harbour black-bellied sandgrouses, because this field work was not exclusively focused in sampling this species, and because most locations were so intensively sampled that there is little room for any geographical bias.

To assess whether the current absence (or the scarcity) of a breeding population from the rest of the eastern islands island can be explained by differences with regards

to Fuerteventura in particular habitat variables, we also performed 594 line transects in Lanzarote and 77 in La Graciosa during 2005 in the semidesert areas of that we deemed more suitable for the species.

For each bird heard or seen, the perpendicular distance to the observer's trajectory was estimated (overflying birds were disregarded) to later obtain estimates of detectability and density (Buckland *et al.* 2008). Previous training with a laser range-finder helped to reduce inter-observer variability in distance estimates.

We measured habitat characteristics by averaging three visual estimations on 25-m radius circular plots located at 125, 250 and 375 m along the line transect. These variables were: coverage of grass, annual forbs, shrubs, and trees (all of them in percentages); mean height of the shrub layer (in cm); rocky cover (in percentage); soil typology (according to the following classes: 0-lava fields, 1-stone/gravel soils, 2-compact soils, 3-sandy soils, and 4-loose sand dunes); and altitude above sea level (measured with GPS receptors). The amount of any agricultural land-use (in percentage) was estimated in two 250-m width bands on both sides of the transects. Other variables were measured on 1:25,000 maps: the distances from the centre of each transect to the nearest paved road (in m) and the nearest city (in m); the length of paved roads (in m) and dirt tracks (in m) within circles of radius 250-m centred on each transect; and the maximum slope terrain in a circle of radius 250-m. Finally, we also used a normalized difference vegetation index (NDVI, range 0-255) as a radiometric index of photosynthetic activity (the larger the value, the more vigorous vegetation). Raw data were ten-day synthesis at 1 km² spatial resolution obtained from the sensor VEGETATION onboard the SPOT satellite (available freely at <http://free.vgt.vito.be/>). We built monthly maximum composite of NDVI images, averaging from 1999 to 2004 (cloudy pixels were assigned a value of zero in the ten-day images, so they were never

selected as the monthly maximum used to build the composite). These values were assigned afterwards to the transects.

The transects were grouped in 18 strata in Fuerteventura, 11 in Lanzarote and just 1 in La Graciosa according to their habitat characteristics and geographical proximity (Figure 1). The Table 1 summarizes the range and mean values for these variables in both islands and Table 2 shows the sampling effort and area covered by each stratum.

Population size estimates

To estimate population sizes we first obtained an estimate of the density in each stratum. In order to do this, we used distance sampling methods, first building a model for the detectability of the species, and then considering the actual counts adjusted for this previous model (Thomas *et al.* 2002). For calculating the detection model, the detection distances (i.e., the perpendicular distances from the transect line at which birds were detected) were right-truncated, thus excluding outliers as recommended by Buckland *et al.* (2001). Then six models were fitted, all of them commonly used to explain the loss of detectability as a function of the distance from the transect line (the further the distance, the lower the probability of detecting a given individual), and the respective probabilities of detection within strips of width equal to the truncated distance were estimated. We calculated a global detection probability function, applied to every stratum. Models were evaluated according to AICc and given weights as: $W_i = \exp(-0.5\Delta AICc) / \sum \exp(-0.5\Delta AICc)$ (Burnham and Anderson 2002). Detectability models were built with Distance 5.0 software (Thomas *et al.* 2004).

We estimated the confidence intervals for the abundance applying a randomization procedure. First, we generated 2,000 random values of probability of detection which lay within the confidence intervals given by Distance. Then, we generated the same amount of random bootstraps of the transects within each stratum to estimate the average number of sandgrouses recorded per 500 m transects (Davison and Hinkley 1997). The density in each trial was calculated according to the probability of detection randomly assigned to that trial (and considering the truncation distance). Finally, we took the 90% confidence intervals for the abundance within each stratum using the bias-corrected and accelerated version of the bootstrap (BCa, DiCiccio and Efron 1996). To obtain the abundance estimate for the whole island the whole sample of all transects was bootstrapped proportionally to the surface of each stratum (weighting each transect in each stratum according to the balance between the proportion of area covered by that stratum and the proportion of transects made on it). This randomization procedure was carried out in Microsoft Excel using PopTools 3.0 (<http://www.cse.csiro.au/poptools/>).

The population size of the species in each stratum was calculated multiplying the estimated densities (mean, lower and upper 90% confidence intervals) by its area. When the lower end of the confidence interval was lower than the actual number of individuals detected, we substituted it by this last amount.

Habitat-relationships models

Species occurrence (absence=0; presence=1) in the sample of 0.5 km line-transects in Fuerteventura was modelled with the 15 original descriptors as explanatory variables, and analyzed using classification trees with Statistica 6.0 (StatSoft 2001). This is a statistical tool where the response variable undergoes successive univariate

splits, according to threshold values of the explanatory variables that maximize the differences between the two resulting groups of samples. Classification trees deal with nonlinear relationships between response and explanatory variables, and with interactions among the latter, and thus are suitable for modelling complex ecological scenarios (Venables and Ripley 1999; De' Ath and Fabricius 2000). We applied misclassification costs according to the detectability of the species, to account for false negatives (the recorded absence in 500 m transects not being always true). In order to do that, we defined misclassification costs as 100 for the presence of the species and 55 for the absence (i.e., proportional to the detectability). To classify the samples in the absent or present groups we used a threshold equal to the prevalence of the species in the whole sample of transects (7.9%). The predictive power of the obtained classification tree was evaluated by means of a cross-validation procedure using 20 four-fold random sampling iterations.

To test the adequacy of Lanzarote and La Graciosa to the Black-bellied Sandgrouse, the classification tree model built for Fuerteventura was applied to the line transects made in these islands. Likewise, the correlation between the predicted probabilities of occurrence (from the tree model) and the estimated abundance (from distance sampling) in Fuerteventura was applied to strata in Lanzarote and La Graciosa to predict the abundance that could potentially reach the grouse.

We also tested for seasonal differences in habitat preferences. In order to do so, we first chose the non-urban transects that were sampled in both March and July and selected those in which we detected any Black-bellied Sandgrouse. Then we performed a multivariate analysis of variance on this reduced set of data, including the season (March, July) as a factor and the same environmental predictors used in the

classification tree as response variables (except percent cover of agricultural land, and cover and height of trees, which were zero for all these transects).

Results

Detectability models and abundance estimates

We did a total of 1,864 line transects during the study period in the three islands in which we registered 156 contacts with 436 individuals (only at Fuerteventura). After visual inspection of the data, the perpendicular distance was truncated at 130 m, thus excluding nine contacts with a total of 29 birds.

The best detectability model was the half-normal key function ($W_i = 0.23$), although the others were fairly sensible alternatives ($\Delta AICc$ lower than 1.80, see Table 3). In addition, all of them provided reasonable fits according to the Cramer-von Mises goodness-of-fit test and the diagnostic plots. Confidence intervals for both the probability of detection and the effective strip width were wide, ranging from 30% to 52% of the mean value. The weighted average probability of detection within 130 m was 0.63, resulting in an average effective strip width (ESW) of 81.7 m at each side of the transect line (95% CI, 67.9 – 99.1).

The areas with the highest densities recorded were Cotillo-Majanicho, Jandía jable, Tefía-Ampuyenta and Triquivijate, with more than seven sandgrouses per km^2 (Table 2). Sandgrouse abundance was highly and significantly correlated with the frequency of occurrence of the species in the 30 sampling sectors of Fuerteventura, Lanzarote and La Graciosa ($r=0.958$, $p<0.001$; $\text{birds} / \text{km}^2 = -41.96 \times \ln [1 - \text{frequency}]$; data from Table 2).

The average number of sandgrouses per sampling stratum ranged from zero individuals in Jandía mountains, Montaña Lengua, Tetir-Rosario and the volcanic

bedrocks, to more than 300 individuals in Jandía jable, Tefía-Ampuyenta, Tindaya and Triquivijate (Table 4). These four areas include only 16.7% of surface of Fuerteventura (251 km²) but 70% of the whole sandgrouse population. The estimate of total population size was 2,906 (90% CI: 2,363 – 3,562) for the whole Fuerteventura island. No birds were registered in nearby Lanzarote or La Graciosa.

Habitat relationships in the breeding season

The classification tree was highly significant ($\chi^2 = 122.5$, d.f. = 10, $p < 0.001$, Figure 2), with a correct prediction of presence-absence of the species in 68.6% of all the transects, and 87.1% of transects where the species was present. The 20 cross-validations showed relatively high correct classifications of the whole sample (61.8%; sd = 2.2%) or the sub-sample of the 93 transects where the species was present (70.9%; sd = 3.6%), that were significantly different from the null hypothesis (50%; $p < 0.001$ in both t-tests).

The most important variables in habitat preferences of Black-bellied Sandgrouse were those related to lithology. The probability of occurrence increased significantly ($p < 0.05$ in all the splits below mentioned) in locations with a soil typology higher than 2 (compact soils) and a rock cover ranging between 0.3% and 44%. Sandgrouses avoided lava fields and stone/gravel soils with a high rock cover, although they also avoided loose sand dunes or plains without rocks or stones present. Orographic characteristics of the terrain also played a prominent role in the habitat preferences of the species, as the probability of occurrence was higher in areas of relatively flat terrain: less than 27.5% in maximum slope, and at altitudes above 8 m a.s.l. (i.e., avoiding strictly coastal areas). The most important vegetation habitat structure variables were tree and shrub cover: the sandgrouse was only present in completely treeless areas

(cover of trees less than 0.1%) with a shrub cover higher than 0.9%. Vegetation productivity, measured by the NDVI index, determined a probability of occurrence of the sandgrouse many times higher when the NDVI was greater than 53. Finally, human influence on sandgrouse was relatively important, as the species showed a strict avoidance of cultivated fields and places near urban areas (closer than 400 m), and preferred sectors of Fuerteventura with a low density of dirt roads (less than 795 m per 20 ha).

Table 2 shows the predicted frequencies of occurrence of the sandgrouse in transects of 500 m in the Eastern Canary Islands according to the classification tree in Figure 2. The frequencies observed and predicted in the 18 strata of Fuerteventura were significantly correlated ($r= 0.791$, $p<0.001$). Although the species was not detected during our sampling in Lanzarote and La Graciosa in 2005, there are several sectors in these islands where the predicted frequencies of occurrence are relatively high in comparison with the frequencies observed in Fuerteventura. The sectors potentially more adequate to the Black-bellied Sandgrouse are La Graciosa, Rubicón, Teguíse and Guatiza, where the predicted frequency of occurrence of the species was higher than the prevalence of the sandgrouse in Fuerteventura. According to the equation relating density to frequency of occurrence, sandgrouse could hypothetically reach densities as high as 4-6 birds / km² in the above mentioned sectors of Lanzarote and La Graciosa.

Seasonal differences in habitat relationships

A total of 602 line non-urban transects were sampled in both March and July. Within those, we detected any Black-bellied Sandgrouse in 49 transects in March and 27 in August. Habitat use was significantly different between months (MANOVA, $F_{11,64}=2.05$, $p=0.039$), although seasonal differences in habitat preferences explained a

low amount of variance (25.9%; Wilk's Lambda=0.741). The distance to urban settlements was the only variable with significant seasonal differences (*a posteriori* one-way ANOVA: $F_{1,74}=7.07$, $p=0.010$). Sandgrouses were closer to human settlements in summer than in March (Figure 3).

The number of birds detected did not significantly vary between seasons (Wilcoxon matched pairs test: $Z=0.533$, $p=0.594$) and the number of birds recorded in the 16 sampled strata were significantly related (Spearman rank order correlation: $r_s=0.573$, $p=0.020$). The only exception to this common pattern of spatial variation in sandgrouse numbers was Lajares-Oliva and Triquivijate, which showed a notable reduction from March to summer.

Discussion

Population size and distribution

The estimated population of the Black-bellied Sandgrouse in the Canary islands is considerably larger than previously reported (Emmerson 1999; BirdLife-International 2004). Several sources of bias could explain the discrepancies between the 2,906 birds reported in this study and the 700-1,200 sandgrouses estimated previously for the Canaries. First, previous counts did not consider the likely of underestimation due to detectability problems. Second, we made a very large sampling effort trying to cover the entire island of Fuerteventura, while previous field work was restricted only to the presumed better areas for the species in this island (350 km² in Emmerson 1999). Third, our survey was carried out at the beginning of the breeding season (March), while previous ones were made in the winter season (December-January) or in the driest part of the summer (July) when the birds may be more mobile and seemingly more difficult to census because they aggregate in large flocks. Thus, we suggest our estimate as the

reference guide for future comparisons on the size of the breeding population of Black-bellied sandgrouses in Fuerteventura.

Four geographic areas include a large part of the population (70%) in a relatively small area of the island (17% of the island area). Of these, three are largely within Special Protection Areas classified under Birds Directive and could benefit from the protection regime established there (these are Jandía jable [63% within SPA], Triquivijate [59%] and Tindaya [74%], that together group half of the island population). Contrastingly, the Tefia-Ampuyenta area, first in absolute number of individuals and in population density, is mostly unprotected (11% within SPA). Therefore, this important area should be considered in future conservation programs of the regional Canary Government, especially considering that it is also an important area for other bird species of conservation concern (Cream-Coloured Courser: Carrascal *et al.* 2007; Houbara bustard: Carrascal *et al.* 2008).

Habitat relationships: biotic vs. abiotic features

The pattern of habitat relationships found for the species in Fuerteventura largely agree with those reported in continental populations, in particular the fact that its abundance correlates negatively with the extent of agricultural land in active use (Suárez *et al.* 1997; but see Cardoso *et al.* 2007). Abiotic features of the landscape had a much greater relevance for the Black-bellied Sandgrouse in this study, which shows that there is room for habitat selection even in apparently simple habitats in terms of human uses and vegetation structure. For example, the species mainly occurs in flat plains of compact soils with an intermediate abundance of stones, while both bare bedrocks and loose sand dunes are avoided. This preference pattern may be advantageous against predation, first because the stones (and short shrubs) could help to conceal clutches

(Lloyd *et al.* 2000; Cardoso *et al.* 2007; Znari *et al.* 2008), and second because flat grounds –as opposed to undulating terrains– may help to spot predators (Ferns and Hinsley 1995). The low dependence of the species on vegetation structure is somewhat unexpected given its seed-feeding habits and the use they make of shrubby vegetation when foraging at high temperatures (Hinsley 1994; Suárez *et al.* 1999; Mian 2003). This habitat relationship probably show the high capacity of the Black-bellied Sandgrouse to exploit food and the thermal patchiness of habitat under constraining environmental conditions (Hinsley *et al.* 1993; Hinsley 1994).

Overall, the habitat relationships of the sandgrouse do not substantially change between seasons (nor does the abundance pattern), which is in agreement with the studies in continental populations (Martínez *et al.* 1998; but see Cardoso *et al.* 2007). Nonetheless, we identified a seasonal change linked to anthropic disturbance: tolerance to nearby cities and villages is higher during the summer than during the breeding period. The explanation for this seasonal change in habitat use might be the intense use of artificial water supplies, which are mostly located in the village surroundings and attract the flocks during the stringent aridity of summer months (Knight 1989).

It is important to stress that the density of roads has a detrimental effect on the sandgrouse, as has been reported for this species in continental studies (Cardoso *et al.* 2007) and for a number of other steppe-land birds in the western Palearctic (Little Bustard: Silva *et al.* 2004; Houbara bustard: Carrascal *et al.* 2006; Cream-Coloured Courser: Palomino *et al.* 2008). This result is particularly worrying if we take into account that infrastructure and urban development has been identified as a new threat for birds in areas that kept until recently a large proportion of well conserved habitats (such is the case of Spain: Madroño *et al.* 2005). The Canaries are most vulnerable to

this form of land transformation due to their small sizes and the current pressure to devote the land to recreational facilities.

Sandgrouses in Lanzarote: why are they not there?

Black-bellied sandgrouses *Pterocles orientalis orientalis* are potent fliers that could seemingly cross the narrow sea stretch between Fuerteventura and Lanzarote (13 km) or perhaps towards the northwest coast of Africa (100 km), where the same subspecies exists and shows some degree of nomadism (see references and maps in Birdguides 2004). Lanzarote and Fuerteventura are similar regarding the environmental characteristics that successfully depict the habitat preference of the species, and there is indeed evidence of a former unquantified breeding population in Lanzarote (Martín and Lorenzo 2001). On the same vein, our models predict a medium density of Black-bellied Sandgrouses in several areas of Lanzarote, most notably for those in which some sparse observations have recently been recorded (Emmerson and Lorenzo 2007). Then, the current absence of a breeding population of Black-bellied Sandgrouses in Lanzarote is intriguing.

Lanzarote has less water resources than Fuerteventura, in particular less natural or artificial ponds for cattle that sandgrouses use as watering-places (pers. obs. and Polatzek 1909 in Emmerson 1999). Thus, the scarcity of watering-places may constrain the potential range within the island, because the availability of water has been identified as a main factor responsible for the presence of the species (Cardoso *et al.* 2007). Also, the present environmental conditions in the islands derive from the past history of human disturbance by agriculture and grazing, which could partly explain the absence of the species in Lanzarote. For example, in a study of biotic and abiotic factors that limited the range size of the endemic Canary Islands stonechat (*Saxicola dacotiae*),

Illera et al. (2006) found that habitat structure and food availability at fine scales differed between the otherwise similar islands of Lanzarote and Fuerteventura, being more favourable in the latter. The authors attribute these differences to the lasting effects of longer and stronger human impacts (via clearance of shrubs and erosion) in Lanzarote (see also Achord *et al.* 2003 for another example on lasting effects of previous human impacts). Similar fine-grained effects may be hindering the settlements of sandgrouses in this island.

Acknowledgements

This paper was funded by projects CGL2005-02642/BOS of the Spanish Ministry of Educación y Ciencia and by a CENTINELA for the monitoring and management of Macaronesian endangered species (Interreg III-B Açores-Canarias-Madeira 2000-2006). We also thank Claire Jasinski for improving the English.

References

- Achord, S., Levin, P. S. and Zabel, R. W. (2003) Density-dependent mortality in Pacific salmon: the ghost of impacts past? *Ecol. Lett.* 6: 335-342.
- Bibby, C. J., Burgess, N. D., Hill, D. A. and Mustoe, S. H. (2000) *Bird Census Techniques*. London: Academic Press.
- Birdguides, L. (2004) *Birds of the Western Palearctic interactive*. Oxford, UK: Oxford University Press.
- BirdLife-International (2000) *Threatened birds of the world*. Barcelona and Cambridge, UK: Lynx Editions and BirdLife International.
- BirdLife-International (2004) *Birds in Europe. Population estimates, trends and conservation status*. London: BirdLife International.

- Buckland, S. T., Anderson, D. R., Burnham, K. P., Laake, J. L., Borchers, D. L. and Thomas, L. (2001) *Introduction to distance sampling*. Oxford: Oxford University Press.
- Buckland, S. T., Anderson, D. R., Burnham, K. P., Laake, J. L., Borchers, D. L. and Thomas, L. (2004) *Advanced distance sampling*. Oxford: Oxford University Press.
- Buckland, S. T., Marsden, S. and Green, R. E. (2008) Estimating bird abundance: making methods work. *Bird Conserv. Internatn.* 18: S91-S108.
- Burnham, K. P. and Anderson, D. R. (2002) *Model selection and multimodel inference. A practical information-theoretic approach*. New-York: Springer-Verlag.
- Cardoso, A. C., Poeiras, A. S. and Carrapato, C. (2007) Factors responsible for the presence and distribution of black bellied sandgrouse *Pterocles orientalis* in the Nature Park "Vale do Guadiana". *Ardeola* 54: 205-215.
- Carrascal, L. M., Seoane, J., Palomino, D. and Alonso, C. L. (2006) Preferencias de hábitat, estima y tendencias poblacionales de la avutarda hubara *Chlamydotis undulata* en Lanzarote y la Graciosa (Islas Canarias). *Ardeola* 53: 251-269.
- Carrascal, L. M., Seoane, J., Palomino, D. and Alonso, C. L. (2007) *El corredor sahariano en España. I Censo Nacional (2005-2006)*. Madrid: SEO/BirdLife.
- Carrascal, L. M., Palomino, D., Seoane, J. and Alonso, C. L. (2008) Habitat use and population density of the houbara bustard *Chlamydotis undulata* in Fuerteventura (Canary Islands). *African J. Ecol.* 46: 291-302.
- Davison, A. C. and Hinkley, D. V. (1997) *Bootstrap methods and their application*. Cambridge: Cambridge University Press.
- De' Ath, G. and Fabricius, K. E. (2000) Classification and regression trees: a powerful yet simple technique for ecological data analysis. *Ecology* 81: 3178-3192.

- De Borbón, M. N., Barros, C. and De Juana, E. (1999) El gregarismo en las gangas ibérica y ortega. Pp. 195-213 in Herranz, J., Suárez, F., eds. *La ganga ibérica (Pterocles alchata) y la ganga ortega (Pterocles orientalis) en España*. Madrid, Spain: Ministerio de Medio Ambiente
- De Juana, E. (1997) Family Pteroclididae (Sandgrouse). Pp. 30-59 in Del Hoyo, J., Elliot, A., Sargatal, J., eds. *Handbook of the Birds of the World. Volume 4. Sandgrouse to Cuckoos*. Barcelona: Lynx Edicions
- DiCiccio, T. J. and Efron, B. (1996) Bootstrap confidence intervals (with discussion). *Statistical Science* 11.
- Emmerson, K. (1999) Distribución y abundancia en las islas Canarias. Pp. 109-115 in Herranz, J., Suárez, F., eds. *La ganga ibérica (Pterocles alchata) y la ganga ortega (Pterocles orientalis) en España*. Madrid, Spain: Ministerio de Medio Ambiente
- Emmerson, K. and Lorenzo, J. A. (2007) Ganga Ortega *Pterocles orientalis*. Pp. 258-261 in Lorenzo, J. A., ed. *Atlas de las aves nidificantes en el archipiélago canario (1997-2003)*. Madrid: Dirección General de Conservación de la Naturaleza-Sociedad Española de Ornitología
- Fernández-Palacios, J. M. and Martín Esquivel, J. L., (eds). 2001 *Naturaleza de las Islas Canarias. Ecología y Conservación*. Santa Cruz de Tenerife: Turquesa.
- Ferns, P. N. and Hinsley, S. A. (1995) Importance of topography in the selection of drinking sites by sandgrouse. *Funct. Ecol.* 9: 371-375.
- Groombridge, B., (ed). 1992 *Global biodiversity: status of the Earth's living resources*. London: Chapman and Hall.
- Hinsley, S. A., Ferns, P. N., Thomas, D. H. and Pinshow, B. (1993) Black-Bellied Sandgrouse (*Pterocles orientalis*) and Pin-Tailed Sandgrouse (*Pterocles*

- alchata*) - Closely related species with differing bioenergetic adaptations to arid zones. *Phys. Zool.* 66: 20-42.
- Hinsley, S. A. (1994) Daily time budgets and activity patterns of sandgrouse (Pteroclididae) in contrasting arid habitats in Spain and Israel. *J. Arid Environm.* 26: 373-382.
- Illera, J. C., Diaz, M. and Nogales, M. (2006) Ecological traits influence the current distribution and range of an island endemic bird. *J. Biogeog.* 33: 1192-1201.
- Knight, M. H. (1989) Importance of borehole water to doves and sandgrouse in the semi-arid southern Kalahari. *South African J. Wild. Res.* 19: 42-46.
- Lloyd, P., Plaganyi, E., Lepage, D., Little, R. M. and Crowe, T. M. (2000) Nest-site selection, egg pigmentation and clutch predation in the ground-nesting Namaqua Sandgrouse *Pterocles namaqua*. *Ibis* 142: 123-131.
- Madroño, A., González, C. and Atienza, J. C., (eds). 2005 *Libro Rojo de las aves de España*. Madrid: Dirección General de Conservación de la Naturaleza-SEO/BirdLife.
- Martín, A. and Lorenzo, J. A. (2001) *Aves del archipiélago canario*. La Laguna, Tenerife: Francisco Lemus.
- Martínez, C., Suárez, F., Yanes, M. and Herranz, J. (1998) Distribución y abundancia de la ganga ibérica *Pterocles alchata* y la ganga ortega *Pterocles orientalis* en España. *Ardeola* 45: 11-20.
- Mian, A. (2003) On Biology of Houbara Bustard (*Chlamydotis undulata macqueenii*) in Balochistan, Pakistan: Influence of Vegetative Characters on Distribution. *OnLine Journal of Biological Science* 3: 484-495.
- Onrubia, A. and Andrés, T. (2005) Impact of human activities on steppic-land birds: a review in the context of the Western Palearctic. Pp. 185-209 in Bota, G.,

- Morales, M. B., Mañosa, S., Camprodon, J., eds. *Ecology and conservation of steppe-land birds*. Barcelona: Lynx Edicions & Centre Tecnològic Forestal de Catalunya
- Palomino, D., Seoane, J., Carrascal, L. M. and Alonso, C. L. (2008) Competing effects of topographic, lithological, vegetation structure and human impact in the habitat preferences of the Cream-coloured Courser. *J. Arid Environm.* 72: 401-410.
- Rodríguez, O., García, A. and Reyes, J. A. (2000) Estudio fitosociológico de la vegetación actual de Fuerteventura (Islas Canarias). *Vieraea* 28: 61-98.
- Silva, J. P., Pinto, M. and Palmeirim, J. M. (2004) Managing landscapes for the little bustard *Tetrax tetrax*: lessons from the study of winter habitat selection. *Biol. Conserv.* 117: 521-528.
- StatSoft, I. (2001). *STATISTICA (data analysis software system)*. StatSoft, Inc., Tulsa, Oklahoma.
- Suárez, F., Martínez, C., Herranz, J. and Yanes, M. (1997) Conservation status and farmland requirements of pin-tailed sandgrouse *Pterocles alchata* and black-bellied sandgrouse *Pterocles orientalis* in Spain. *Biol. Conserv.* 82: 73-80.
- Suárez, F., Hervás, I., Levassor, C. and Casado, M. A. (1999) La alimentación de la ganga ibérica y la ganga ortega. Pp. 215-229 in Herranz, J., Suárez, F., eds. *La ganga ibérica (Pterocles alchata) y la ganga ortega (Pterocles orientalis) en España*. Madrid, Spain: Ministerio de Medio Ambiente
- Suárez, F. and Herranz, J. (2004) Ganga Ortega *Pterocles orientalis*. Pp. 265-269 in Madroño, A., González, C., Atienza, J. C., eds. *Libro rojo de las aves de España*. Madrid: Dirección General para la Biodiversidad-SEO/BirdLife

- Suárez, F., Hervás, I., Herranz, J. and Del Moral, J. C. (2006) *La ganga ibérica y la ganga ortega en España: población en 2005 y método de censo*. Madrid: SEO/BirdLife.
- Thomas, L., Buckland, S. T., Burnham, K. P., Anderson, D. R., Laake, J. L., Borchers, D. L. and Strindberg, S. (2002) Distance sampling. Pp. 554-552 in El-Shaarawi, A. H., Piegorisch, W. W., eds. *Encyclopedia of Environmetrics*. Chichester: John Wiley & Sons
- Thomas, L., Laake, J. L., Strindberg, S., Marques, F. F. C., Buckland, S. T., Borchers, D. L., Anderson, D. R., Burnham, K. P., Hedley, S. L., Pollard, J. H. and Bishop, J. R. B. (2004). *Distance 5.0, Release Beta 2*. Research Unit for Wildlife Population Assessment, St. Andrews.
- Venables, W. N. and Ripley, B. D. (1999) *Modern Applied Statistics with S-PLUS*. New York: Springer-Verlag.
- Znari, M., Aourir, M., Radi, M. and Melin, J.-M. (2008) Breeding biology of the Black-bellied Sandgrouse *Pterocles orientalis* in west-central Morocco. *Ostrich* 79: 53-60.

	FUERTEVENTURA				LANZAROTE				LA GRACIOSA			
	mean	sd	min	max	mean	sd	min	max	mean	sd	min	max
ALTITUDE	147.1	107.5	1	584	141.7	101.6	2	481	42.2	21.4	11	96
SLOPE	11.5	12.0	0	73	7.0	6.9	0	42	8.8	7.8	1	37
ROCK COVER	27.8	24.6	0	100	25.4	29.6	0	100	15.0	23.7	0	90
SOIL INDEX	2.4	0.9	0	4	2.2	1.1	0	4	3.0	0.9	1	4
DMIN-URBAN	2766.2	2097.0	0	14828	1877.2	1285.2	0	7966	2615.0	1490.6	44	5371
DMIN-ROAD	1240.7	940.4	0	5400	950.4	956.9	0	6000	5588.6	1471.3	2929	8572
L-TRACKS	204.5	288.0	0	1850	321.9	298.1	0	1210	186.7	239.2	0	900
L-ROADS	46.2	141.1	0	1100	68.8	161.7	0	1000	0.0	0.0	0	0
GRASS COVER	3.6	7.6	0	68	5.0	10.4	0	75	0.0	0.4	0	3
FORBS COVER	11.4	11.8	0	70	12.7	13.6	0	100	5.3	8.4	0	56
SHRUB COVER	9.5	7.9	0	57	6.2	6.3	0	45	14.6	7.7	4	42
H-SHRUB	0.3	0.3	0	2	0.2	0.2	0	1	0.3	0.1	0	1
TREE COVER	0.5	2.5	0	34	0.0	0.0	0	0	0.0	0.0	0	0
NDVI	61.0	9.5	30	103	59.9	10.7	39	110	58.7	8.7	48	90
AGRIC. COVER	1.0	6.9	0	90	8.7	21.4	0	100	0.2	1.7	0	15

Table 1. Environmental characteristics (mean \pm standard deviation and range) of the sampled areas in Lanzarote and Fuerteventura (eastern Canary Islands). See Figure 1 for their respective location. Number of 0.5-km transects is 1,184 for Fuerteventura and 594 for Lanzarote. ALTITUDE: mean altitude above sea level (in m); SLOPE: average slope of the terrain (in %); ROCK COVER: cover of rocks and stones (in %); SOIL INDEX: index size of soil grain (0: volcanic soils; 1: stony soils; 2: compact sandy soils; 3: sandy soils; 4: loose dunes); DMIN-URBAN: minimum distance to the nearest city (in m); DMIN-ROAD: minimum distance to the nearest paved road (in m); L-TRACKS: length of unpaved tracks (in m) per 20 ha; L-ROADS: length of paved roads (in m) per 20 ha; FORBS COVER: cover of forbs (in %); GRASS COVER: cover of grass (in %); SHRUB COVER: cover of shrubs (in %; mostly chamaephytes and small phanerophytes of genus *Suaeda*, *Salsola*, *Launaea*, *Lycium* and *Euphorbia*); H-SHRUB: mean height of them shrubs (in cm). TREE COVER: cover of trees (in %; mainly *Tamarix canariensis* and *Phoenix canariensis*). NDVI: normalized difference vegetation index; AGRIC. COVER: cover with agricultural uses (in %).

Stratum	Effort	Area (km ²)	Observed birds/km ²	Observed frequency	Predicted frequency	Predicted birds/km ²
Fuerteventura						
Betancuria	114	331	0.18	0.01	0.01	0.42
Castillo Sur	46	47	1.74	0.04	0.11	4.89
Corralejo	69	19	0.00	0.00	0.08	3.50
Cotillo-Majanicho	26	12	7.88	0.12	0.09	3.96
Fimapaire-Finimoy	80	50	3.12	0.05	0.13	5.84
Jandía (Jable)	103	68	7.54	0.18	0.09	3.96
Jandía (Mountains)	41	100	0.00	0.00	0.00	0.00
Lajares-Oliva	63	34	6.46	0.19	0.12	5.36
Malpaíses Norte	72	94	0.00	0.00	0.01	0.42
Malpaíses Sur	14	43	0.00	0.00	0.00	0.00
Montaña Lengua	37	20	0.00	0.00	0.04	1.71
Morro Jable	17	20	3.02	0.06	0.07	3.05
Tefía-Ampuyenta	85	75	8.68	0.13	0.11	4.89
Tetir-Puerto del Rosario	60	163	0.00	0.00	0.03	1.28
Tindaya	106	52	7.01	0.17	0.16	7.32
Triquivijate	106	57	8.47	0.18	0.16	7.32
Tuineje	91	116	1.38	0.03	0.04	1.71
Vigán-Giniginámar	63	200	0.17	0.02	0.01	0.42
Lanzarote						
Rubicón	95	54	0.00	0.00	0.11	4.89
Teguisse	36	26	0.00	0.00	0.10	4.42
Guatiza	98	55	0.00	0.00	0.09	3.96
Tías-Puerto del Carmen	53	48	0.00	0.00	0.07	3.05
Famara	94	59	0.00	0.00	0.06	2.60
Soo-Tinajo	31	35	0.00	0.00	0.06	2.60
Zonzamas	32	29	0.00	0.00	0.03	1.28
Haría	27	69	0.00	0.00	0.02	0.85
Playa Quemada	25	11	0.00	0.00	0.01	0.42
Geria-Tiagua	34	46	0.00	0.00	0.00	0.00
Malpaíses	69	161	0.00	0.00	0.00	0.00
La Graciosa	77	18.7	0.00	0.00	0.14	6.47

Table 2. Sampling effort per stratum (Effort: in number of 0.5-km transects) and summary of sampling results. For each stratum, it is given the area (km², excluding urban areas from strata in Figure 1) along with the estimated density of individuals (number of birds/km²), the observed frequency (proportion of transects in which the species was detected) and the predicted frequency according to the classification tree model. Finally, the predicted density is given, according to the linear relationship between the density and the natural logarithm of one minus the frequency (that is, the observed frequency in Fuerteventura and the predicted frequency for Lanzarote and La Graciosa: $r=0.943$, $n=18$ strata in Fuerteventura, $\text{birds/km}^2 = -41.96 \cdot \ln[1 - \text{frequency}]$).

	$\Delta AICc$	AICc	W	P	ESW
Half-normal (cosine)	0	1480.2	0.233	0.60 (0.51, 0.69)	77.4 (66.9, 89.6)
Half-normal (polynomial)	0.013	1480.2	0.231	0.63 (0.54, 0.74)	82.3 (70.0, 96.7)
Hazard-rate (cosine)	0.692	1480.9	0.165	0.61 (0.47, 0.79)	79.3 (61.6, 102.1)
Negative exponential (polynomial)	0.832	1481.1	0.153	0.61 (0.47, 0.79)	79.6 (61.3, 103.3)
Negative exponential (cosine)	1.266	1481.5	0.124	0.66 (0.53, 0.82)	85.8 (68.8, 107.4)
Hazard-rate (polynomial)	1.798	1482.0	0.095	0.63 (0.66, 0.81)	93.4 (86.2, 105.4)
Weighted average				0.63 (0.52, 0.76)	81.7 (67.9, 99.1)

Table 3. Models fitted to the detection distances truncated at 130 m (n= 156 contacts with 436 individuals), ordered increasingly according to their Akaike's Information Criterion corrected for small sizes (AICc) values (i.e., from larger to smaller reliability). W is the weight given to each model according to the formula $W_i = \exp(-0.5\Delta AICc) / \sum \exp(-0.5\Delta AICc)$ (Burnhman and Anderson, 2002). It is also given the detection probability within 130 m and its 95% confidence interval (P), and the effective strip width ESW (note that $ESW = P \times 130$, allowing for rounding errors). The Cramer-von Mises goodness of fit test, which measures the difference between the empirical distribution function and the probability distribution function in a quantile-quantile plot, was non-significant in every model (Buckland *et al.* 2004).

Geographic area	Abundance	Cumulative population (%)	Cumulative area (%)
Tefia-Ampuyenta	640 (367-1097)	22	5
Jandia jable	524 (343-814)	40	10
Triquivijate	494 (311-776)	58	13
Tindaya	369 (252-549)	70	17
Lajares-Oliva	191 (98-402)	77	19
Tuineje	164 (66-386)	83	27
Fimapaire-Finimoy	160 (60-347)	88	30
Cotillo-Majanicho	98 (35-243)	92	31
Castillo Sur	83 (20-269)	95	34
Morro Jable	59 (8-214)	97	35
Betancuria	59 (2*-325)	99	57
Vigan-Giniginar	35 (1*-223)	100	70
Corralejo	0 (0-0)	100	71
Jandia macizo	0 (0-0)	100	78
Malpaises Norte	0 (0-0)	100	84
Malpaises Sur	0 (0-0)	100	87
Montana Lengua	0 (0-0)	100	89
Tetir-Rosario	0 (0-0)	100	100
Fuerteventura	2906 (2363-3562)		

Table 4. Abundance (with 90% confidence interval) of Black-bellied Sandgrouses in the geographic areas considered. Lower-level confidence intervals marked with asterisks were estimated to be zero but are substituted here for the actual number of birds detected. No birds were registered in either Lanzarote or la Graciosa and thus we estimate that there is no current breeding population in these islands.

Figure 1. Location of a) the study area, b) the geographical strata (light lines) and the centres of line transects (dots) in Lanzarote, La Graciosa (above) and Fuerteventura (below) c) samples where Black-bellied sandgrouses were detected. Areas that could not be surveyed are in grey.

Figure 2. Classification tree describing the pattern of habitat preferences of the Black-bellied Sandgrouse in Fuerteventura (Eastern Canary Islands). The probability of presence of the species is expressed below each box as a percentage. Bold-lined boxes indicate final environmental conditions. The number of transects meeting the previous set of conditions is shown inside each box. The splitting variables and threshold values selected refer to left branches of the tree, so that right branches met opposite conditions. See Table 1 for the acronyms of the variables.

Figure 3. Minimum distance to urban settlements of 0.5 km transects where Black-bellied Sandgrouse was present in spring (n=49) and summer (n=27) in Fuerteventura working with the same sample of 602 transects common to both periods. It is also shown the average minimum distance to the nearest city of all transect centres (in m).

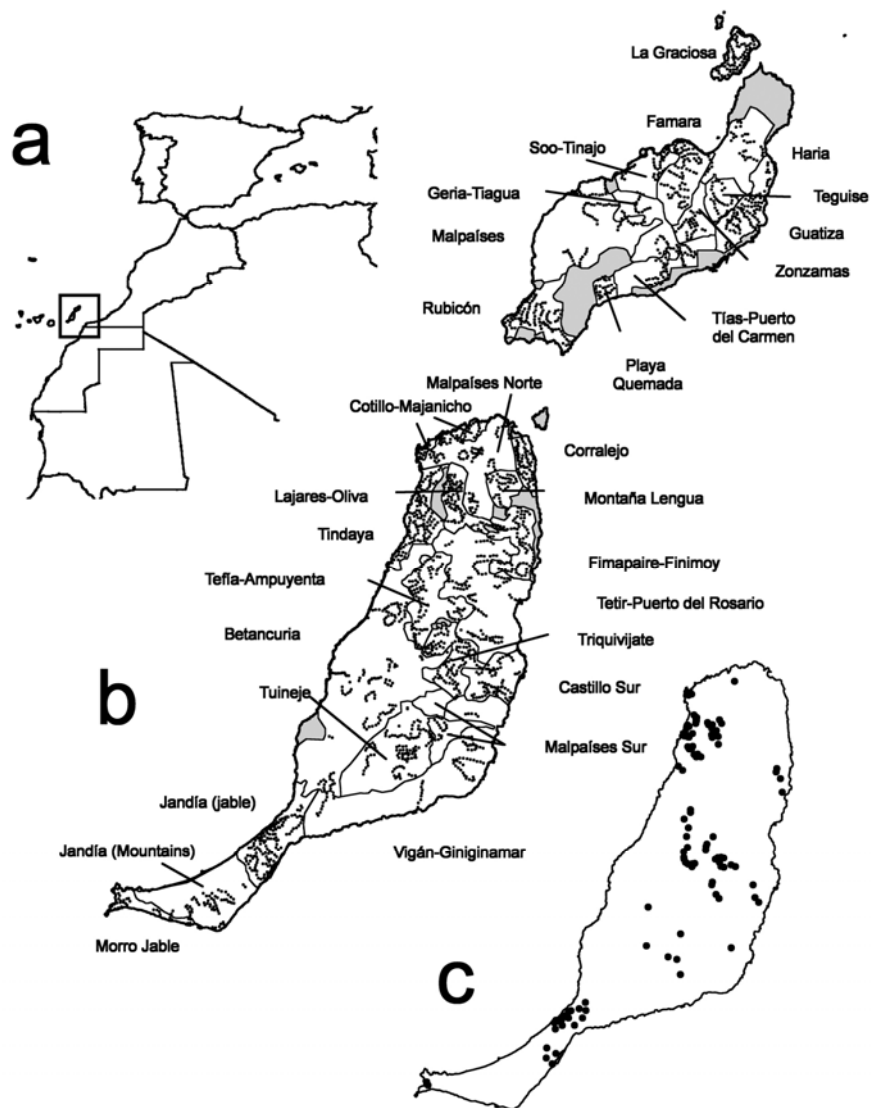


Figure 1

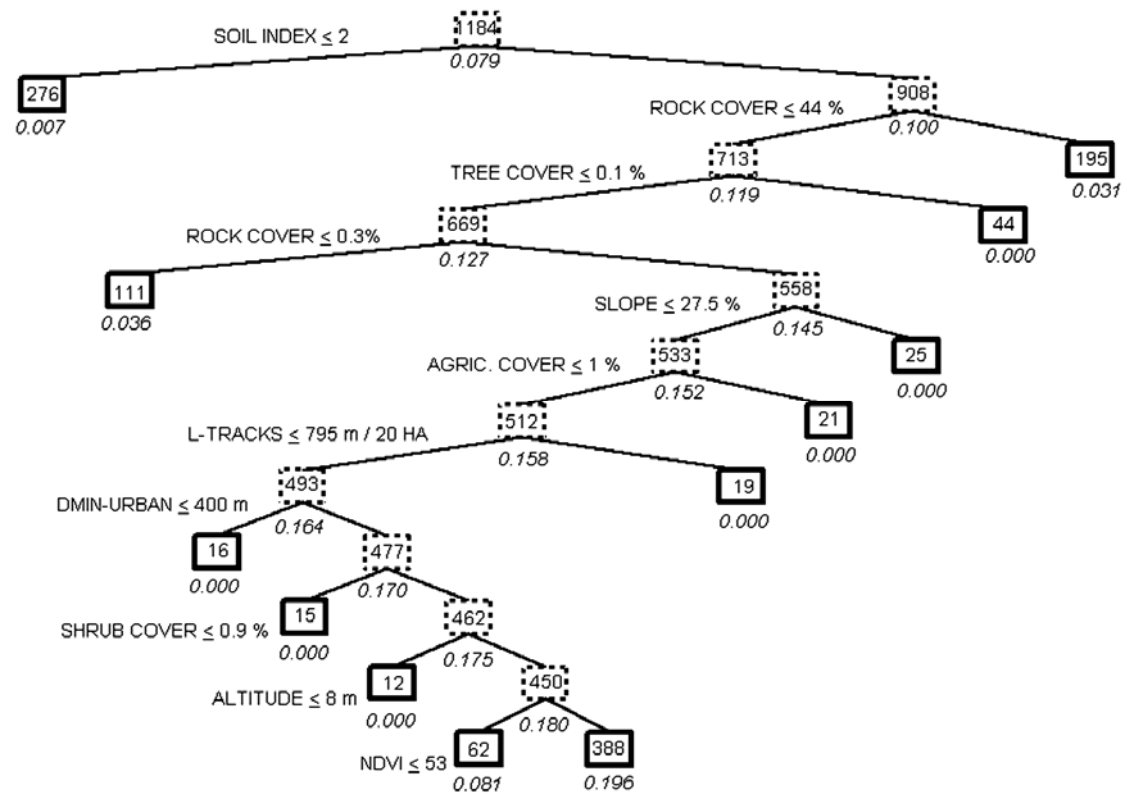


Figure 2

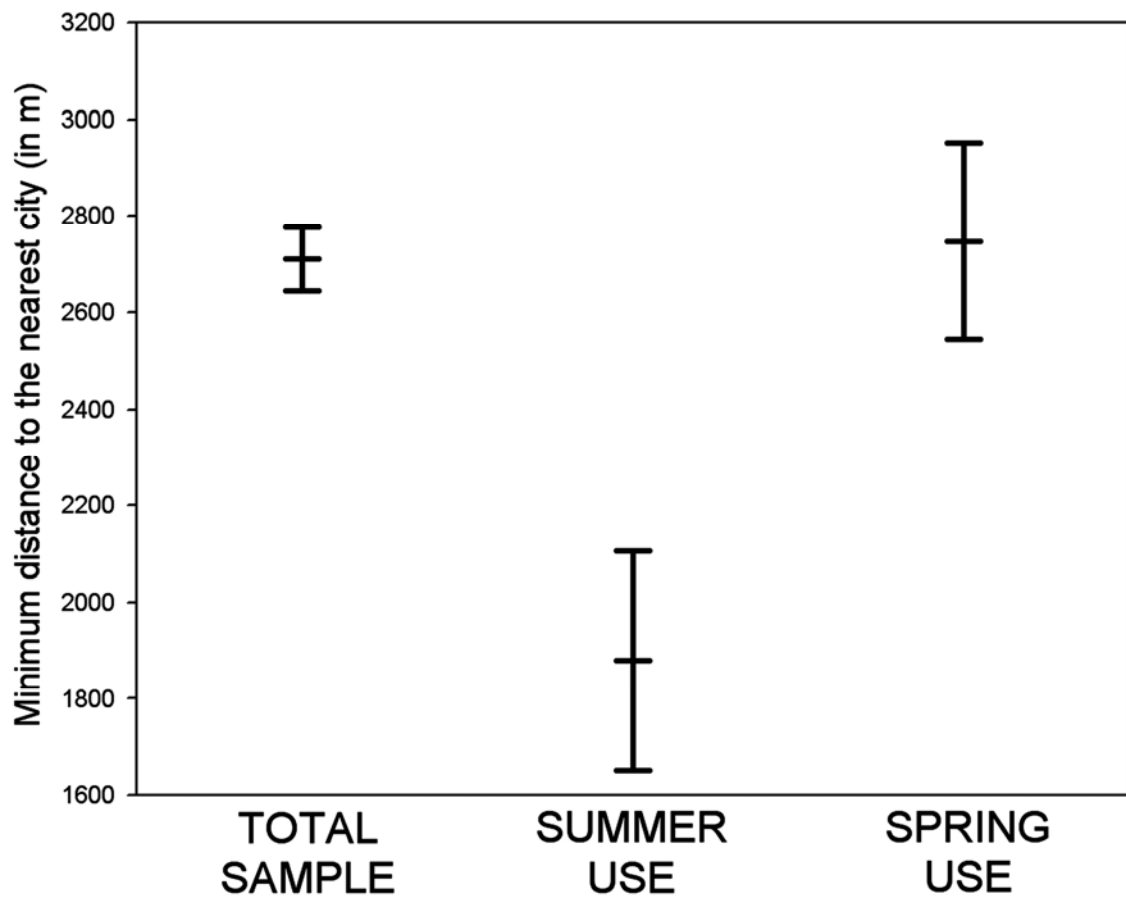


Figure 3